

1 RH: Sustainable harvest in space

2

3

4 THE EFFECT OF HARVESTING ON THE SPATIAL
5 SYNCHRONY OF POPULATION FLUCTUATIONS

6

7

8 Steinar Engen

9 Department of Mathematical Sciences, Centre for Biodiversity Dynamics, Norwegian
10 University of Science and Technology, N-7491 Trondheim, Norway

11

12 Francisco J. Cao

13 Departamento de Física Atómica, Molecular y Nuclear, Universidad Complutense de Madrid, Avenida
14 Complutense s/n, 28040 Madrid, Spain.

15

16 Bernt-Erik Sæther

17 Department of Biology, Centre for Biodiversity Dynamics, Norwegian University of Science
18 and Technology, N-7491-Trondheim, Norway

19

20

21 *Corresponding author:* Bernt-Erik Sæther, Email: Bernt-Erik.Sather@ntnu.no, Phone:

22 +4790578544

23

24

25

26

27

28

29 ABSTRACT

30 Harvesting in space affects, in general, the spatial scale of the synchrony in the population
31 fluctuations, which determines the size of the areas subjected to simultaneous quasi-
32 extinction risk. Here we show that harvesting reduces the population synchrony scale if it
33 depends more strongly on population fluctuations than the density dependence of the growth
34 rate in the absence of harvesting. We show that constant and proportional harvesting always
35 increases the spatial scale, using a theta-logistic model for density regulation. We also
36 provide exact scaling results under harvesting for the Beverton-Holt and the Ricker stock-
37 recruitment models that are commonly applied, e.g. in fisheries. Our results indicate that
38 harvest in areas with large abundances should be encouraged to avoid increase of the spatial
39 scale of synchrony in the population fluctuations that can lead to unexpected quasi-extinction
40 of populations over large areas. Our results quantify this harvesting impact giving the
41 resulting scales of spatial synchrony of population synchrony. This emphasizes the
42 importance of estimating the form of density dependence as well as the dependency of
43 harvest upon population density of exploited populations, in order to get reliable predictions
44 of the size of areas that can undergo simultaneous quasi-extinction.

45

46 *Key words: spatial dynamics, harvest theory, stochastic dynamics, density dependence,*
47 *environmental stochasticity.*

48

49 1. Introduction

50

51 One of the greatest contributions of harvest theory has been to provide general
52 principles for how populations of exploited species should be managed. For example, the
53 concept of Maximum Sustainable Yield denotes the maximum equilibrium harvest rate
54 (Reed, 1978; Quinn and Deriso, 1999), which in a stable environment is obtained by keeping
55 the population at the size that maximizes the population growth rate. For a logistic model of
56 density regulation this population size is half the carrying capacity K . Another example of
57 such a general principle is threshold harvesting, providing the optimal strategy for reduced
58 risk of extinction or the probability of reduction in population size to an unacceptable low
59 level in a stochastic environment (Lande et al., 1995, 1997). Although these principles have
60 been difficult to implement in practice, they still provide important guidelines for
61 identification of overall goals for the management strategy of exploited species.

62 Traditionally harvesting theory has focused on modelling the dynamics of a single
63 population over time. This ignores that many intensively harvested species are spread out in
64 space and migrate over large areas. Accordingly, a central aim in obtaining a sustainable
65 management strategy is therefore to consider the effects of harvesting at relative small
66 geographical scales and then distribute the total quotas among sub-areas within the
67 distributional range of the species according to some set of predetermined criteria (Lavsund
68 et al, 2003; Cadgan et al., 2017). An additional complication introduced by space is that
69 fishermen and hunters are often selective in their choice of areas to harvest. This may occur
70 because of logistic constraints to access areas, management regulations (e.g. formation of
71 reserves) or may be based on some kind of spatial harvesting strategy (e.g. by specifically
72 visiting areas with high yields) (Ling and Milner-Gulland, 2008). Over the last decades there
73 has therefore been increased focus on how harvest will affect interactions among sub-

74 populations localized in different areas (McCullough, 1996; Jonzén et al., 2001; Fogarty and
75 Botsford, 2007; Ling and Milner-Gulland, 2008; Kerr et al., 2010; Williams and Hastings,
76 2013; Burgess et al., 2014; Benson et al., 2015; Takashina and Mougi, 2015). This implies
77 that the dynamics at a single location cannot be understood without including spatial aspects
78 of the dynamics. A nice illustration has been the recent attention to formation of marine
79 reserves as a management option (Tuck and Possingham, 2000; Gell and Roberts, 2003).
80 General management principles for exploited species should therefore include a spatial
81 dimension (Hilborn et al., 2003; Schindler et al., 2010; Hilborn, 2012; Benson et al., 2015).

82 Spatial synchrony is a concept usually referring to the temporal covariance or
83 correlation between two quantities, for example population densities, measured at different
84 spatial distances (Ranta et al., 1998; Lande et al., 2003; Liebhold et al., 2004). Ignoring
85 sampling variances the correlations are by definition one at distance zero, and in a large
86 homogeneous area they often decrease monotonically towards zero at large distances where
87 fluctuations are independent (Liebhold et al., 2004; Sæther et al. 2007) . The spatial scale of
88 the synchrony is a distance where correlations typically are considered to be large below that
89 value and small at larger distances. In an isotropic model a circle with radius equal to the
90 scale then typically is an area with rather similar population densities. Such scales can be
91 defined in many different ways, for example where the correlation passes 0.5, $e^{-1} \simeq 0.37$ or
92 0.10 (Bjørnstad and Falck, 2001), but can also be expressed through the shape of the whole
93 autocorrelation function (Engen et al., 2005). Lande et al. (1999) scaled the correlation (or
94 covariance) function by a constant factor to produce a distribution and used the standard
95 deviation of this distribution as a measure of spatial scale. If the correlation has an
96 exponential form, this scale is then where the correlation equals $e^{-1} \simeq 0.37$, whereas for a
97 Gaussian form it equals $e^{-0.5} \simeq 0.61$.

98 The spatial scale of the synchrony in population fluctuations is an important
99 characteristic with important ecological ramifications to describe how population dynamics
100 covary in space. For instance, synchronous populations tend to be more vulnerable to
101 extinction or quasi-extinctions than populations that show spatially asynchronous population
102 fluctuations (Heino et al., 1997; Heino, 1998; Palmqvist and Lundberg, 1998), and if the
103 spatial scale is large, extinctions may occur surprisingly fast over large areas (Engen 2007).
104 Importantly, the degree of spatial scaling of the population dynamics can also be understood
105 from basic processes affecting fluctuations in population size. This is illustrated by the model
106 of Lande et al. (1999), who analyzed a spatial population model with spatially correlated
107 noise, local density regulation and dispersal. Using the scale defined as standard deviation
108 their main result is expressed by the equation

$$109 \quad l_N^2 = l_e^2 + \frac{ml_f^2}{\gamma} . \quad (1)$$

110 Here, l_N is the spatial scale of the spatial autocovariance function for population densities, l_e
111 is the corresponding scale for the spatial correlation of the environmental noise, m is the
112 dispersal rate and l_f is the standard deviation of dispersal distance, while γ is the strength of
113 local density regulation. In models with no spatial dimension $1/\gamma$ is the mean return time to
114 equilibrium (May, 1974). An important conclusion from this analysis is that even the effect of
115 small and moderate dispersal over small distances may be very large if the strength of local
116 density dependence is small. A small γ makes effects of local migration spread out in space
117 through generations, creating large areas with similar densities. If there is no dispersal then
118 $l_N = l_e$, which is Moran's (1953) classical result.

119 The purpose of the present paper is to analyze how different harvesting strategies
120 affect the spatial scale of population fluctuations. We show that harvesting may either

121 increase or decrease the spatial scale, depending on spatial characteristics of the harvesting
122 strategy, that is, the harvest at different locations with different population densities.

123

124 **2. Dynamic population model in time and space**

125

126 A general class of population models in continuous time with no spatial structure has the
127 form

$$128 \quad dN = [r - g(N)]Ndt + \sigma_e NdB(t), \quad (2)$$

129 where $g(N)$ is an increasing function of population size N , expressing density regulation
130 through how it reduces the growth rate $r-g(N)$ as population size increases. The parameter σ_e^2
131 is the environmental noise and $dB(t)$ is the infinitesimal increment of a standard Brownian
132 motion during time dt so that the expectation $E [dB(t)]=0$ and $\text{var} [dB(t)]=dt$. The temporal
133 variance of the noise term $N\sigma_e dB$ is accordingly $N^2\sigma_e^2 dt$. The carrying capacity K is defined
134 as the equilibrium population size of the corresponding deterministic model so that $r -$
135 $g(K)=0$. A harvesting strategy can be defined through a harvesting rate $h(N)$, meaning that
136 $h(N)dt$ individuals are harvested during dt . The dynamics then change to

$$137 \quad dN = [r - g(N)]Ndt - h(N)dt + \sigma_e NdB. \quad (3)$$

138 A central focus for the harvest theory is to find choices of harvesting rates $h(N)$ that
139 produce large mean annual yields while maintaining long-term sustainability of the harvest
140 (Hilborn and Walters, 1992; Quinn and Deriso; 1999). A general conclusion is that constant
141 harvesting in a fluctuating environment leads to unstable dynamics with large risk of
142 extinction (Beddington and May, 1977). As another extreme, thresholds harvesting reduces
143 the risk of unexpected severe reductions in population size and hence represents the most
144 sustainable approach when the population size is known (Reed, 1978, 1979; Lande et al.

145 1995, 1997), although temporal fluctuations in the yield will be large (Lande et al., 1997;
 146 Sæther et al., 2001).

147 In absence of harvesting, Lande et al. (1999) analyzed the spatial dynamics of $N(x, t)$,
 148 the population density at location $x = (x_1, x_2)$ in the plane at time t , using the linearization at
 149 the carrying capacity which at any location and time replaces $[r - g(N)]N$ by
 150 $-Kg'(K)(N - K) = -\gamma(N - K)$, where γ is the strength of density regulation. A logistic
 151 model is defined by $g(N) = rN/K$ giving $\gamma = r$. Adding density independent dispersal at rate m
 152 to this model and writing $f(y)$ for the two-dimensional distribution of dispersal distance y for
 153 dispersing individuals, the spatio-temporal model in the neighborhood of K takes the simple
 154 linear form

$$155 \quad dN(x, t) = -\gamma[N(x, t) - K]dt + m dt \int [N(x - y, t) - N(x, t)]f(y)dy + \sigma_e K dB(x, t), \quad (4)$$

156 where the integration is performed over the entire two-dimensional plane and the noise is
 157 approximated by the noise at $N = K$. The noise terms are here spatially correlated with spatial
 158 autocorrelation $\rho_e(y) = E[dB(x, t)dB(x + y, t)] / dt$. This model yields stationary fluctuations
 159 of population densities with a spatial correlation $\rho_N(y) = \text{corr}[N(x, t), N(x + y, t)]$.

160 Using this model, Lande et al. (1999) derived their simple result for the spatial scale
 161 of fluctuations in densities given by Eq. 1. Defining the scale along a given direction in
 162 space, which without loss of generality may be chosen as the first axis y_1 , as the standard
 163 deviation of the scaled autocorrelation function, so that the squared scale is

$$164 \quad l_N^2 = \int y_1^2 \rho_N(y_1, y_2) dy_1 dy_2 / \int \rho_N(y_1, y_2) dy_1 dy_2, \text{ they derived Eq. 1, where all scales refer to the}$$

165 same direction in space.

166

167 **3. The effect of harvesting**

168 *3.1 Theta-logistic dynamics*

169

170 Adding harvesting as in Eq. 3 at each point in space now yields exactly the same kind
171 of linearized model as in Eq. 4, but with a smaller deterministic equilibrium K_h and a
172 different strength of density regulation γ_h , because harvesting by itself is a density regulating
173 process (Fig. 1). The spatial scale under harvesting is accordingly found simply by replacing
174 γ by γ_h in Eq. 1. Harvesting increases the spatial scale if $\gamma_h < \gamma$, and will otherwise reduce
175 it. The new parameters under harvesting may in general be calculated numerically for any
176 choice of harvesting strategy characterized by a new deterministic equilibrium K_h .

177 A wide array of models has been proposed to describe the form of density regulation
178 in natural populations. Here we will apply the theta-logistic model $g(N) = r(N/K)^\theta$ (Gilpin
179 and Ayala, 1973), which defines a general class of models by varying just one parameter θ .
180 Large values of θ indicate strong density regulation around K , whereas for small θ density
181 regulation starts to act at small population sizes (Lande et al., 2003). Another advantage of
182 the theta-logistic model is that it provides for specific values of θ commonly used models for
183 the form of the density-regulation such as the logistic ($\theta=1$) as in Fig. 1, the loglinear ($\theta=0$)
184 model using appropriate scaling of the growth rate, as well as the ceiling model ($\theta = \text{infinite}$).
185 In Appendix A we derive a simple analytic solution for γ_h under theta-logistic dynamics for a
186 given choice of K_h , and harvesting rates expressed as power functions $h(N) = cN^{\psi+1}$. This
187 represents constant harvest for $\psi = -1$ and proportional harvesting for $\psi = 0$. Choosing
188 $\psi = \theta$ the density dependence generated by harvesting has the same form as the density
189 regulation under no harvesting (Fig. 1). A large positive value of ψ corresponds to threshold
190 harvesting with no harvesting below K_h and strong harvesting as soon as the density
191 exceeds K_h . For this model it is shown in Appendix A that

192
$$\gamma_h = \gamma + r(\psi - \theta)[1 - (K_h / K)^\theta]. \quad (5)$$

193 In combination with Eq. 1 this shows that the spatial scale of population fluctuations
 194 increases by harvesting if the strategy is given by the parameter ψ smaller than θ . This
 195 conclusion is general in the sense that it holds for any values of the parameter r , K and θ
 196 describing the dynamics under no harvesting.

197 Many types of density regulation can be approximated by a theta-logistic model using
 198 the first order expansion of $\ln g(N)$ at $N=K$ writing in general $\theta = d \ln g(N) / d \ln N$
 199 evaluated at $N=K$. A range of harvesting strategies with small reduction of the carrying
 200 capacity can be expressed by first order expansions of $\ln[h(N) / N]$ writing
 201 $d \ln[h(N) / N] / d \ln N = \psi$ at $N=K$, so that Eq. 5 then can serve as an approximation (see
 202 Appendix A for details).

203 A general conclusion emerging from Eq. 5 is that γ_h is smaller than γ if $\psi < \theta$.
 204 Accordingly, if the harvesting is done so that it by itself represents a weaker density
 205 regulation than that given by θ in the absence of harvesting, then from Eq. 1 the spatial scale
 206 of density fluctuations increases by harvesting (Fig. 1). In the opposite case for stronger
 207 harvest than the density-regulation, the spatial scale decreases and approaches the scaling of
 208 the spatial autocorrelation in the environment l_e producing the Moran (1953) -effect (Fig.
 209 2).

210 Threshold harvesting corresponds to the limit obtained as ψ approaches infinity. Let
 211 us keep K_h constant, meaning that one always harvest down to the threshold K_h at any
 212 location if the density is larger than K_h and otherwise do not harvest at those locations. From
 213 Eq. 5 it appears that γ_h takes large values when ψ increases giving approximately $l_N = l_e$.
 214 However, for very large values of ψ the approximation obtained by linearization at K_h will
 215 not be accurate. Nevertheless, the effect of dispersal will disappear approximately with

216 spatial scale corresponding approximately to Moran's (1953) result for strategies close to
217 threshold harvesting.

218 We use these general results for the theta-logistic model to make a simple illustration
219 depicted in Fig. 1 of the effects of constant and proportional harvesting in the logistic model
220 choosing $K_h = 0.75K$. The spatial scale under harvesting is given by Eq. 1 replacing γ with
221 minus the slope γ_h at $N = K_h$. From Eq. 5 constant harvesting gives $\gamma_h / \gamma = 1/2$ and
222 proportional harvesting $\gamma_h / \gamma = 0.75$. Thus, constant and proportional harvesting make
223 $\gamma_h < \gamma$ and hence increases the spatial scale of the synchrony in the population fluctuations.
224 In contrast, harvesting proportional to N^2 gives $\gamma_h / \gamma = 1$. Hence, the spatial scaling
225 remains unchanged if the harvesting has the same form as the original density regulation, that
226 is, proportional to N^2 for the logistic model. Finally, stronger dependence on N (here
227 exemplified by N^3 that gives $\gamma_h / \gamma = 5/4$) reduces the spatial scale of the population
228 dynamics.

229

230 *4.2 Special cases: the Beverton-Holt and the Ricker stock recruitment models*

231

232 The two most commonly used models in fishery biology are the stock-recruitment
233 models of Ricker (1954) and Beverton and Holt (1957). Assume that the fecundity at small
234 densities is f and that adult survival s is unaffected by density. The Beverton-Holt and Ricker
235 models then describe the fecundity (recruitment) at density N by the relations $f / (1 + \alpha N)$
236 and $fe^{-\beta N}$, respectively, where α and β are positive parameters expressing the strength of
237 the density regulation. The corresponding population growth dN in continuous time is
238 accordingly $N[f / (1 + \alpha N) + s - 1]dt$ and $N[fe^{-\beta N} + s - 1]dt$. We show in Appendix B that θ
239 equals $1 / (1 + \alpha K)$ and $\beta K / (e^{\beta K} - 1)$, which always take values between 0 and 1 for both

240 models. Hence, using the theta-logistic approximation constant and proportional harvesting
241 provide $\gamma_h < \gamma$ for both models, and these harvesting strategies result in increased spatial
242 scale of population fluctuation. On the other hand, harvesting rate proportional to N^2 ($\psi =$
243 1) will always decrease the spatial scale.

244 In Appendix B we also compute γ_h exactly for the two stock-recruitment models. This
245 enables us to compare how the effect of a given harvest (e.g. from K to $K/2$) on the spatial
246 scaling depends on variation in vital rates (Fig. 3). For both the Beverton-Holt (Fig. 3a) and
247 the Ricker (Fig. 3b) model the spatial scale in the population dynamics increases with
248 decreasing adult fecundity rate. However, the scaling effects of increasing adult survival rate
249 differed between the two models. In the Ricker-model the scaling effect of the harvest was
250 relatively unaffected or decreased with increasing adult survival (Fig. 3a), whereas the
251 increase in spatial scaling of a given harvest was particularly large among long-lived species
252 subject to a Beverton-Holt model of density regulation (Fig. 3b). The accuracy of the theta-
253 logistic approximation for the two models (Appendix B) was good as long as the carrying
254 capacity under harvesting was larger than about 90 % of K (Fig. A1). The approximation is
255 seen to be conservative in the sense that it overestimates the value of $1/\gamma_h$ and hence the
256 effects of harvest on the spatial scaling of the population dynamics.

257

258 5. Discussion

259

260 Here we propose that management of spatially structured exploited populations living
261 in a fluctuating environment should encourage harvesting representing a stronger type of
262 density regulation than that in the absence of harvesting (Fig. 1) to avoid that harvesting
263 increases the spatial synchrony in the population fluctuations (Fig. 2). For the theta-logistic
264 model of density regulation (Gilpin and Ayala, 1973) we provide an analytical result (Eq. 5)

265 for the critical strength of density regulation from harvesting that must be exceeded to
266 prevent an increase in the synchrony of the population dynamics over larger areas. Such an
267 increase of the scale should be avoided to reduce the risk of extinction or reduction in
268 population size to a low level in large areas (Foley, 1997; Heino et al., 1997; Heino, 1998;
269 Palmqvist and Lundberg, 1998; Ciannelli et al., 2013). This extends previous conclusions
270 from temporal models of optimal harvesting in stochastic environments (Lande et al., 1995,
271 1997; Sæther et al., 2001) indicating that sustainability requires a strong dependency of
272 harvest on population size, being small for small population sizes and large when the
273 population density exceed some threshold. Thus, the general results obtained here can be
274 used as foundation for developing more specific harvest strategies that include spatial
275 variation in harvest rates, which has been identified as a general problem in management of
276 exploited species (Hilborn et al., 2003; Schindler et al., 2010; Hilborn 2012).

277 Here we have analyzed the effects of harvesting on the spatial dynamics by assuming
278 harvesting at a rate $cN^{\psi+1}$ at location x and time t . For a closed population with appropriate
279 choice of c this covers a wide array of harvesting strategies, including constant ($\psi = -1$) and
280 proportional ($\psi = 0$) harvesting, as well as approaching pure threshold harvesting in the limit
281 as ψ approaches infinity. For the theta-logistic model of density regulation (Gilpin and
282 Ayala, 1973; Lande et al., 2003) our results show that harvesting will increase the spatial
283 scale if performed in a way corresponding to choosing $\psi < \theta$. In practice, θ varies among
284 species but in many taxa it often is located between 0 and 1 (Sæther and Engen, 2002; Sæther
285 et al., 2002). For example, Lillegård et al. (2005) estimated θ for Norwegian spring spawning
286 herring *Clupea harengus* to be about 0.57. Furthermore, the commonly applied stock-
287 recruitment models in fisheries, the Beverton-Holt and Ricker models, has parameters
288 $\theta = d \ln g(N) / \ln N |_{N=K}$ between zero and 1 when approximated by a theta-logistic model.
289 However, in many populations of terrestrial vertebrates θ may be larger than 1 (Sæther and

290 Engen, 2002; Sæther et al., 2002, 2008) so that proportional harvesting is always expected to
291 increase the spatial scale and harvesting strategies with stronger dependence on population
292 size (larger values of ψ) should be encouraged.

293 We have derived a general result for the effects of harvesting on the spatial synchrony
294 of the population fluctuations for the theta-logistic model expressed by γ_h given by Eq. 5 in
295 combination with equation Eq. 1 (Fig. 2). For more general types of models, such as the one
296 obtained by replacing N^θ with a general form of density regulation $g(N)$, it is difficult to
297 find transparent analytical results. However, the theta-logistic is a very general type of model
298 that covers a large number of different forms of density regulation (Lande et al, 2003) and for
299 conservative harvesting only down to about 90% of K (Lande et al., 1997), it works well as
300 an approximation (see Appendix A). Furthermore, for the commonly applied Beverton-Holt
301 and Ricker types of density regulation we have provided exact results for the effects of
302 harvesting expressed by power functions (Appendix B). These results are, however, not as
303 general as those for the theta-logistic model because an increase or decrease in spatial scale
304 by harvesting depend on all parameters in the model (and not only θ as in the theta-logistic
305 model).

306 These analyses are based on several simplifying assumptions that may restrict their
307 practical applicability. Firstly, we assume that the parameters characterizing the population
308 dynamics are exactly known and that there is no biases in the population estimates. For
309 instance, in non-spatial models uncertainty in estimates of population sizes should imply
310 cautious harvest strategy by increasing the threshold based on the estimated population size
311 as an insurance against overestimation of population size that can lead to too aggressive and
312 non-sustainable harvesting (Engen et al., 1997). Secondly, the analyses are based on a
313 spatially homogeneous model. This means that this model is inappropriate for examining the
314 effects of refuges, e.g. caused by establishment of non-harvest reserves (Gell and Roberts,

315 2003) or the presence of logistically inaccessible areas (Ling and Milner-Gulland, 2008).
316 Thirdly, we assume no effects of fluctuations in the age-structure, which often strongly affect
317 temporal variation in population size of many species subject to harvest (Quinn and Deriso,
318 1999; Solberg et al., 1999).

319 Although these simplifying assumptions mean that it may be difficult to apply our
320 proposed strategies in practice, the analysis may still serve as a general guideline for
321 harvesting spatially distributed populations. For instance, our results imply that one should
322 encourage searching for high density areas for harvesting and avoid harvesting in areas with
323 small densities as this not only gives larger mean annual yield, but also decreases the spatial
324 scale of population fluctuations. Following such guidelines will tend to break down the
325 synchrony so that unexpected severe reductions of population size or even quasi-extinctions
326 over large areas can be avoided. In practice is not possible to perform harvesting
327 proportional to a given $h(N)$ at any location in space, but our results can still be applied to
328 give general suggestions for managers determining quotas. To utilize the theory it is
329 important to have good knowledge about the form of density regulation as well as estimates
330 of densities in space at each discrete time when quotas are determined, which requires
331 extensive sampling (Benson et al., 2015; Cadigan et al., 2017) as well as spatial statistical
332 methods for smoothing (Cressie and Wikle, 2011). From a given estimated spatial field of
333 densities one can locate areas with for example large, medium and small densities (or some
334 finer class division), and give large quotas relative to total population in the large density
335 areas, much smaller in the medium, and no harvesting in the small density area, with details
336 depending on the form of density dependence. The density dependence, estimate of K as well
337 as mean densities in the three areas may suggest harvesting efforts ensuring no increase in
338 spatial scale of population synchrony.

339 In this paper we are not addressing the problem of optimal harvesting in terms of
340 mean annual yield. Based on analyses of non-spatial models it is known that the optimal
341 solution is threshold harvesting provided known population sizes (Lande et al., 1997), and
342 strategies close to that when population sizes are estimated with some uncertainty (Engen et
343 al., 1997). How to deal with threshold harvesting in space is, however, an extremely complex
344 mathematical challenge. For proportional harvesting the problem was analyzed analytically
345 for small and moderate noise by Engen (2017) and some promising results were obtained.
346 This will without any doubt be an important area for further research.

347

348

349 **Acknowledgments**

350

351 This work was supported by the European Economic Area (EEA) Grants UCM-
352 EEA-ABEL-02-2009 and 005-ABEL-CM-2014A under the NILS -Science and
353 Sustainability Programme, and by the Research Council of Norway through the
354 project SUSTAIN and a Centre of Excellence Grant (SFF-III 223257/F50). FJC
355 acknowledge financial support through grants FIS2010-17440 and FIS2006-
356 05895 of Ministerio de Ciencia e Innovacio'n (Spain), GR35/14-920911, GR35/10-
357 A-920911 and GR58/08-920911 of Universidad Complutense de Madrid and
358 Banco Santander (Spain).

359

360

361 **References**

- 362 Beddington, J. R., May, R.M., 1977. Harvesting natural populations in a randomly fluctuating
363 environment. *Science* 197, 463 -465.
- 364 Benson, A. J., Cox, S. P., Ceary, J. S., 2015. Evaluating the conservation risks of aggregate
365 harvest management in a spatially-structured herring fishery. *Fish. Res.* 167, 101-113.
- 366 Beverton, R. J. H., Holt, S. J., 1957. On the dynamics of exploited fish populations. *Fish.*
367 *Invest. Lond. Ser 2* 19, 1-533.
- 368 Bjørnstad, O. N., Falck, W., 2001. Nonparametric spatial covariance functions: Estimation
369 and testing. *Env. Ecol. Stat.* 8, 53-70.
- 370 Burgess, S. C., Nickols, K. J., Griesemer, C. D., et al., 2014. Beyond connectivity: how
371 empirical methods can quantify population persistence to improve marine protected-
372 area design. *Ecol. Appl.* 24, 257-270.
- 373 Cadigan, N. G., Wade, E., Nielsen, A., 2017. A spatiotemporal model for snow crab
374 (*Chionoecetes opilio*) stock size in the southern Gulf of St. Lawrence. *Can. J. Fish.*
375 *Aq. Sci.* 74, 1808-1820.
- 376 Ciannelli, L., Fisher, J.A.D., Skern-Mauritzen, M., et al. 2013. Theory, consequences and
377 evidence of eroding population spatial structure in harvested marine fishes: a review.
378 *Mar. Ecol. Progr. Ser.* 480, 227-243.
- 379 Cressie, N., Wikle, C.K., 2011. *Statistics for Spatio-temporal Data*. John Wiley & Sons,
380 Hoboken, New Jersey.
- 381 Engen, S. 2007. Stochastic growth and extinction in a spatial geometric Brownian population
382 model with migration and correlated noise. *Math. Biosci.* 209, 240-255.
- 383 Engen, S. 2017. Spatial synchrony and harvesting in fluctuating populations: Relaxing the
384 small noise assumption. *Theor. Pop. Biol.* 116, 18-26.
- 385

386 Engen, S., Lande, R., Sæther, B.-E., 1997. Harvesting strategies for fluctuating populations
387 based on uncertain population estimates. *J. Theor. Biol.* 186, 201 -212.

388 Engen, S., Lande, R., Sæther, B.-E., Bregnballe, T., 2005. Estimating the pattern of
389 synchrony in fluctuating populations. *J. Anim. Ecol.* 74, 601-611.

390 Fogarty, M. J., Botsford, L.W., 2007. Population Connectivity and Spatial Management of
391 Marine Fisheries. *Oceanography* 20, 112-123.

392 Foley, P. 1997. Extinction models for local populations. In: Hanski, I.A., Gilpin, M.E.
393 (Eds.). *Metapopulation biology. Ecology, genetics, and evolution.* Academic Press,
394 San Diego, pp. 215-246.

395 Gell, F. M., Roberts, C.M., 2003. The Fishery Effects of Marine reserves. *World Wildlife*
396 *Fund*, Washington.

397 Gilpin, M. E., Ayala, F.J., 1973. Global models of growth and competition. *Proc. Natl. Acad.*
398 *Sci. USA* 70, 3590 - 3593.

399 Heino, M. 1998. Noise colour, synchrony and extinctions in spatially structured populations.
400 *Oikos* 83, 368-375.

401 Heino, M., Kaitala, V., Ranta, E., Lindström, J., 1997. Synchronous dynamics and rates of
402 extinction in spatially structured populations. *Proc. R. Soc. B Lond.* 264, 481-486.

403 Hilborn, R. 2012. The evolution of quantitative marine fisheries management 1985-2010.
404 *Nat. Res. Mod.* 25, 122-144.

405 Hilborn, R., Quinn, T. P., Schindler, D. E., Rogers, D. E., 2003. Biocomplexity and fisheries
406 sustainability. *Proc. Natl. Acad. Sci. USA* 100, 6564-6568.

407 Hilborn, R., Walters, C. J., 1992. *Quantitative fisheries stock assessment. Choice, dynamics,*
408 *and uncertainty.* Chapman & Hall, New York.

409 Jonzén, N., Lundberg, P., Gårdmark, A. 2001. Harvesting spatially distributed populations.
410 *Wildl. Biol.* 7, 197-203.

- 411 Kerr, L. A., Cadrin, S. X., Secor, D.H. 2010. The role of spatial dynamics in the stability,
412 resilience, and productivity of an estuarine fish population. *Ecol. Appl.* 20, 497-507.
- 413 Lande, R, Engen, S, Sæther, B.-E., 1995. Optimal harvesting of fluctuating populations with
414 a risk of extinction. *Am. Nat.* 145, 728 - 745.
- 415 Lande, R, Engen, S, Sæther, B.-E., 1999. Spatial scale of population synchrony :
416 environmental correlation versus dispersal and density regulation. *Am. Nat.* 154, 271
417 - 281.
- 418 Lande, R, Engen, S, Sæther, B.-E., 2003. *Stochastic Population Dynamics in Ecology and*
419 *Conservation.* Oxford University Press, Oxford.
- 420 Lande, R., Sæther, B.-E., Engen, S., 1997. Threshold harvesting for sustainability of
421 fluctuating resources. *Ecology* 78, 1341 - 1350.
- 422 Lavsund, S., Nygren, T., Solberg, E.J., 2003. Status of moose populations and challenges to
423 moose management in Fennoscandia. *Alces* 39, 109-130.
- 424 Liebhold, A., Koenig, W. D., Bjørnstad, O.N. 2004. Spatial synchrony in population
425 dynamics. *Ann. Rev. Evol. Syst.* 35, 467-490.
- 426 Lillegård, M., Engen, S. Sæther, B.-E., Toresen, R., 2005. Harvesting strategies for
427 Norwegian spring-spawning herring. *Oikos* 110, 567-577.
- 428 Ling, S., Milner-Gulland, E. J., 2008. When does spatial structure matter in models of
429 wildlife harvesting? *J. Appl. Ecol.* 45, 63-71.
- 430 May, R. M. 1974. *Stability and complexity in model ecosystems.* Princeton University Press,
431 Princeton.
- 432 McCullough, D. R. 1996. Spatially structured populations and harvest theory. *J. Wildl.*
433 *Manage.* 60, 1-9.
- 434 Moran, P. A. P. 1953. The statistical analysis of the Canadian lynx cycle. II. Synchronization
435 and meteorology. *Austr. J. Zool.* 1, 291 - 298.

436 Palmqvist, E., Lundberg, P., 1998. Population extinction in correlated environment. *Oikos* 83,
437 359 - 367.

438 Quinn, T. J. I., Deriso, R.B. 1999. Quantitative fish dynamics. Oxford University Press, New
439 York.

440 Ranta, E., Kaitala, V., Lindström, J., 1998. Spatial dynamics of populations. In: Bascompte,
441 J. Solé, R.V. (Eds.). *Modeling Spatiotemporal dynamics in ecology*. Springer -
442 Verlag, Berlin, pp. 47 - 62.

443 Reed, W. J., 1978. The steady state of a stochastic harvesting model. *Math. Biosci.* 41, 273 -
444 307.

445 Reed, W. J., 1979. Optimal escapement levels in stochastic and deterministic harvesting
446 models. *J. Env. Econ. Manage.* 6, 350 - 363.

447 Ricker, W. E. 1954. Stock and recruitment. *J. Fish. Res. Bd. Can.* 11, 559-623.

448 Schindler, D. E., Hilborn, R., Chasco, B. et al., 2010. Population diversity and the portfolio
449 effect in an exploited species. *Nature* 465, 609-612.

450 Solberg, E. J., Sæther, B.-E., Strand, O., Loison, A., 1999. Dynamics of a harvested moose
451 population in a variable environment. *J. Anim. Ecol.* 68, 186 - 204.

452 Sæther, B.-E., Engen, S.. 2002. Pattern of variation in avian population growth rates. *Phil.*
453 *Trans. R. Soc. B Lond.* 357, 1185 - 1195.

454 Sæther, B.-E., Engen, S., Grøtan, V. et al., 2008. Forms of density regulation and (quasi-)
455 stationary distributions of population sizes in birds. *Oikos* 117, 1197-1208.

456 Sæther, B.-E., Engen, S., Grøtan, V. et al., 2007. The extended Moran effect and large-scale
457 synchronous fluctuations in the size of Great tit and Blue tit populations. *J. Anim.*
458 *Ecol.* 76, 315-325.

459 Sæther, B.-E., Engen, S., Matthysen, E., 2002. Demographic characteristics and population
460 dynamical patterns of solitary birds. *Science* 295, 2070 - 2073.

461 Sæther, B.-E., Engen, S., Solberg, E.J., 2001. Optimal harvest of age structured populations
462 of moose *Alces alces* in a fluctuating environment. *Wildl. Biol.* 7, 171 - 179.

463 Takashina, N., Mougí, A., 2015. Maximum sustainable yields from a spatially-explicit
464 harvest model. *J. Theor. Biol.* 383, 87-92.

465 Tuck, G. N., Possingham, H.P., 2000. Marine protected areas for spatially structured
466 exploited stocks. *Mar. Ecol. Progr. Ser.* 192, 89-101.

467 Williams, P. D., Hastings, A., 2013. Stochastic dispersal and population persistence in marine
468 organisms. *Am. Nat.* 182, 271-282.

469

470

471

472

473

474

475 **Legend to figures**

476

477 **Fig. 1.** The mean growth rate $E(dN/dt)$ for the logistic model of density regulation as a
478 function of population density N in the case of no harvesting (solid line) as well as for four
479 different harvest strategies (constant and proportional harvesting as well as harvest
480 proportional to N^2 and N^3) (dashed lines). The mean growth rate at small densities under no
481 harvesting is $r=0.5$ and the carrying capacity without harvesting is $K=1$. The strength of
482 density regulation without harvesting is $\gamma=r=0.5$. All 4 harvesting strategies are scaled by
483 a factor so that the deterministic equilibrium under harvesting is $K_h=0.75K$.

484

485 **Fig. 2.** The spatial scale of the synchrony in population size $l_N = \sqrt{l_e^2 + ml_f^2 / \gamma_h}$, where l_e is
486 the spatial scale for the spatial correlation of the environment noise, m is the dispersal rate, l_f
487 is the standard deviation of dispersal distance and γ_h is the strength of local density
488 regulation under harvesting, as function of the harvesting strategy ψ . The harvesting rate is
489 $cN^{\psi+1}$ with c chosen to give carrying capacity K_h under harvesting. $\psi = -1$ and $\psi = 0$
490 express constant and proportional harvesting, respectively. The dynamics is logistic ($\theta = 1$)
491 and results are shown for 5 different values of K_h/K . The parameter γ_h is found by Eq. 5.
492 Parameter values are $r = 0.3$, $m = 1$, $l_f = 10$, and $l_e = 20$, while K is arbitrary.

493

494 **Fig. 3.** The factor $1/\gamma_h$ (see Eq. 1) determining the spatial scaling of the population
495 dynamics for the Beverton-Holt (a) and Ricker (b) model as function of adult survival for
496 different fecundities f at small densities by proportional harvesting ($\psi = 0$) down to half of
497 the initial carrying capacity, that is $K_h = K/2$. The other parameters are $f=5$ and $s=0.8$.







